

1 **Beaver dams mitigate the impacts of whiplash weather in a fragmented habitat: A Salinas River**
2 **case study**

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11

12 **Abstract**

13 Beaver *Castor canadensis* create ecological refugia against drought, heat stress and fire, and policies to
14 support beaver conservation and recolonization in regions where they have been historically extirpated are
15 increasingly common. Between prolonged periods of drought, arid regions are increasingly challenged by
16 extreme precipitation events that promote flash floods, debris flows, and mudslides—a phenomenon known
17 as “whiplash weather”. Understanding how beaver wetlands respond to whiplash weather will help inform
18 the development of restoration policies targeting the species as a natural climate solution. We used remotely
19 sensed normalized difference vegetation index data to characterize the influence of beaver complexes on
20 riparian greenness dynamics under whiplash weather by comparing three complexes and five nearby
21 reference areas along the Salinas River, California. Our study region is within a remanent patch of the
22 historic range of beaver and is highly impacted by agricultural and urban uses. Despite these limitations to
23 expansion and their low density due to historical extirpation, the Salinas River beaver complexes
24 demonstrated greater riparian greenness resistance to drought and resilience to flood disturbance than the
25 watershed reference areas. Thus, policies supporting beaver re-colonization—even within highly
26 fragmented and anthropogenically impacted habitats—may confer both riparian resistance and resilience to
27 increasingly erratic climatic conditions.

28

29 **Keywords**

30 Beaver *Castor canadensis*; whiplash weather; wetlands; Arid West; restoration; normalized difference
31 vegetation index (NDVI); Salinas River; California

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45 Introduction

46

47 Like many areas globally, California is experiencing increasingly severe and prolonged drought periods
48 caused by anthropogenically driven climate change (Hayhoe et al. 2004; Gutzler and Robbins 2011; Mann
49 and Gleick 2015). Climate change accelerates the drying of plant biomass, extending fire season
50 vulnerability (Keeley and Syphard 2016; Williams et al. 2019; Higuera and Abatzoglou 2021; Swain 2021).
51 Between prolonged periods of drought and heightened wildfire risk, California and other regions of the arid
52 western US now face concerns of too much water—a phenomenon known as “whiplash weather,”
53 characterized by extreme precipitation events following a dry spell that promote flash floods, debris flows,
54 and mudslides (Swain et al. 2018; He and Sheffield 2020; Francis et al. 2022). Given the myriad of socio-
55 environmental consequences to extreme climatic variability, there is a growing need for mitigation and
56 adaptation. Beaver *Castor canadensis* is a keystone species that alters riparian corridors and lakes
57 throughout the boreal and temperate zones; these activities can increase local climate resiliency and
58 resistance to chronic disturbance through their dam building and wetland maintenance activities (Hood and
59 Bayley 2008; Dittbrenner et al. 2018; Brazier et al. 2021; Thompson et al. 2021; Nash et al. 2021). While
60 beaver wetland complexes can support ecological refugia to drought and fire (Hood and Bayley 2008;
61 Fairfax and Small 2018; Fairfax and Whittle 2020), how these wetlands affect ecosystem response to deluge
62 following drought is less well characterized. Documenting beaver wetland complex response to whiplash
63 weather is thus an important consideration in developing robust wetland restoration policies that target
64 beavers as a natural climate solution.

65 Wetland systems are sometimes characterized as the ecological equivalent of a “negative emission
66 technology” due to their ability to store high carbon stocks (Griscom et al. 2017; Windham-Myers et al.
67 2018). Anerobic sediment supports methanogenesis while simultaneously slowing the rate of
68 decomposition, allowing wetlands to store large pools of soil carbon while also supporting improved
69 wildlife habitat quality and water filtration, among other ecosystem services (Page and Dalal 2011). The
70 global extent of wetlands is estimated to be up to 21% of the total global land area although many wetlands

71 have been degraded due to land use practices (Davidson et al. 2018). Some of the largest intact wetland
72 areas occur in North America, primarily Canada and Alaska (Davidson et al. 2018; Tootchi et al. 2019).
73 Historically perceived as undesirable systems (Vileisis 1997), 40-50% of wetlands in the conterminous US
74 were destroyed due to agricultural drainage by the end of the 20th century (Kolka et al. 2021; Fluet-
75 Chouinard et al. 2023). More recently, perception of wetlands has shifted from unwanted to intrinsically
76 valuable due to their many ecosystem services, leading to the development of policies to conserve and
77 restore these systems (Zedler 2000).

78 Beaver create and restore wetlands at minimal cost and are increasingly recognized as an effective
79 natural climate solution (e.g., see California’s 2024 ‘Beaver Bill’ AB 2196; Griscom et al. 2017; Connolly
80 2024). Beavers use tree such as willow, aspen, and cottonwood to build their dams (Dittbrenner et al. 2018;
81 Puttock et al. 2021). Dam establishment slows water flow and creates ponds, increases channel complexity,
82 and facilitates floodplain creation, thus supporting sediment aggradation (Puttock et al. 2021; Grudzinski
83 et al. 2022). Higher surface and groundwater, greater channel complexity comprised of multiple small dams
84 and lateral river extensions, and larger floodplains created by beaver complexes also help to regulate flood
85 events by dampening, storing, and dissipating high flows (Westbrook et al. 2006, 2020; Feiner and Lowry
86 2015). Westbrook et.al (2020) noted that even failed beaver dams slowed downstream flood waves,
87 although beaver can have minimal impact on flood attenuation (Neumayer et al. 2020; Larsen et al. 2021).

88 Regions facing the dual threat of increasing drought and wildfire stress can benefit from beaver
89 recolonization and conservation due to the ecosystem services associated with their dam building activities
90 (Brazier et al. 2021; Thompson et al. 2021). Yet, the fluctuation between two extremes—drought and
91 flooding—may challenge beaver wetland ecosystem services provisioning and it remains uncertain if
92 beaver influenced hydrogeomorphology confers resiliency or resistance to whiplash weather. Resistance
93 is the ability of a system component to withstand change following a perturbation, while resilience is the
94 rate at which a component returns to its reference condition following a perturbation (Pimm 1984).
95 Therefore, systems exhibiting resistance or resilience in response to a perturbation are unlikely to exhibit
96 the other (i.e., a system will not be both resistant and resilient to drought but may be resistant to drought

97 and resilient to flood). Uncertainty in how beaver-influenced areas respond to more extreme weather has
98 raised the call for further investigation at different landscapes and scales (Wohl 2021; Graham et al. 2022).

99 To better understand how beaver activities influence riparian response to whiplash weather in the
100 Arid West, we assessed remotely sensed normalized difference vegetation index (NDVI) patterns in three
101 beaver complexes relative to five comparable areas that did not exhibit evidence of beaver damming
102 activities in the Salinas River watershed in San Luis Obispo County, California (Figure 1). This area is
103 within the boundaries of the historic beaver range but in a remanent patch of its current extent that is highly
104 modified by farming, urbanization, and damming for agriculture (Baker and Hill 2003; Scamardo et al.
105 2022). Beginning in December 2022, an extraordinarily wet season followed a multi-year and increasingly
106 extreme drought period in San Luis Obispo County (National Integrated Drought Information System (U.S.)
107 2011). By characterizing NDVI response during a transition from an average precipitation period to extreme
108 drought conditions to flooding, we explored whether localized beaver impacts on a riparian corridor
109 influence plant community resistance to prolonged drought and heat stress, and whether this sustained
110 resistance correlates with resistance or resilience to a subsequent extraordinary flood event.

111 **Methods**

112

113 **Site Selection**

114 Beavers exert a strong influence on riparian vegetation as a byproduct of their dam building and herbivory.
115 Cutting riparian trees can induce coppicing in some species, and the dam itself elevates water tables,
116 supporting greener vegetation which can be remotely measured using NDVI (Bento et al. 2018). We used
117 NDVI data derived from Landsat imagery that encompassed three known beaver complexes and nearby
118 reference areas along the Salinas River, CA from July 2017 through December 2023 to characterize the
119 effect of direct beaver activities on riparian resiliency after a high peak flow period due to extraordinary
120 rain conditions (December 2022 – March 2023) that followed sustained and increasingly extreme drought
121 conditions (January 2021-November 2022). Our study area is located in San Luis Obispo County, which
122 has a mean annual temperature of 15.4 °C and a mean annual precipitation of 45.6 cm that primarily falls
123 from December through March. The County received 75.3 cm of rain between December 2022 to March
124 2023 (NOAA 2025) causing all known beaver dams along the Salinas to breach with no re-built or newly
125 established beaver dams noted during field visits following the flooding through April 2023.

126 We identified beaver complex locations along the Salinas River in November and early December
127 2022 with field surveys that were supplemented by iNaturalist research-grade observations from 2015 -
128 2022 (GBIF Occurrence Download 2025) and confirmation by local experts who had visited the sites. We
129 focused our study on three beaver complexes and five nearby reference areas; our site selection was limited
130 by public land access points interspersed with private land along a 40 km segment between Atascadero and
131 Paso Robles, CA within the 282 km span of the Salinas River. Although maintained by different beaver
132 families, we grouped some of the dams into complexes because they were directly adjacent to one another
133 in the stream network, making it difficult to distinguish at the aerial scale. We selected the focal sites
134 because they encapsulated riparian areas heavily influenced by beaver presence. Areas that were at least
135 100 meters upstream or downstream of the beaver complexes were haphazardly chosen as reference sites
136 (i.e., areas not directly influenced by beaver damming activities), since beavers generally do not stray more

137 than 50 meters from their complex as they are sluggish on land (Stoffyn-Egli and Willison 2011; Svanholm
138 Pejstrup et al. 2023).

139 The Salinas River is heavily impacted by agricultural water use and is characterized by dry riverbed
140 sections. While beaver influenced stream segments were braided and contained surface water year-round,
141 the reference sites did not have comparable channel complexity and lacked surface water year-round.
142 Cottonwood and willow trees were prominent in the beaver complex areas; reference sites had fewer trees
143 and sparse shrubs in the perimeter (Figure 2). Both beaver and reference sites were constrained by human
144 development along the Salinas River. There was no significant difference in the beaver complex and
145 reference site areas (beaver complex ($56,4571.3 \pm 30,8095.7 \text{ m}^2$) and non-beaver influenced ($45,1562.0 \pm$
146 $16,1301.7 \text{ m}^2$; $p = 0.6$).

147

148 **Study area NDVI patterns**

149 We used ArcGIS Pro to delineate beaver complex and reference sites, extracted Analysis Ready Landsat 8
150 and 9 Images with less than 10% cloud coverage from July 2017 to December 2023 that encompassed the
151 study area, and obtained Band 4 (Red) and Band 5 (near infrared, NIR) .tif files for each image (Data S1-
152 Data S2). Landsat 8 and 9 satellites pass through the same area every 16 days with an 8-day offset. In
153 addition to the acquisition schedule, the constraints on cloud coverage yielded at most three usable images
154 produced per month for each site. We calculated NDVI in a Python notebook as $(\text{NIR} - \text{Red}) / (\text{NIR} + \text{Red})$
155 for each pixel (Drisya et al. 2018). We placed a feature class of 100 random points within each site polygon
156 and calculated the daily NDVI for each point in ArcGIS Pro. We then filtered the NDVI raster set so that
157 any negative values (indicating standing water) were considered null values. For each site and time frame,
158 we used an average of 98.9 of the 100 points to derive the mean NDVI (i.e., non-negative NDVI values).
159 We calculated average daily NDVI within each site which we used to calculate the monthly means of the
160 study sites ($n = 3$ beaver complexes and $n = 5$ reference sites) across the six sub-datasets (Data S3).

161 We subset the NDVI data to account for seasonality (plant community green up versus dry down
162 senescence) and the conditions during the water year as reported by the National Integrated Drought
163 Information System and County of San Luis Obispo Water Resources Advisory Committee Rain and
164 Reservoir Reports (National Integrated Drought Information System (U.S.) 2011; Rain and Reservoir
165 Reports, County of San Luis Obispo 2024). We filtered the data based on local drought status and monthly
166 precipitation records: normal (July 2017 – December 2020, which encompassed periods where some of the
167 region experienced abnormal to moderate drought interspersed with drought-free periods), abnormally dry
168 to extreme drought (January 2021 – November 2022, where 100% of the region was in extreme drought or
169 higher from June 2021 – December 27, 2021), and flood to wet (January 2022 – December 2023). We
170 further filtered the data by site type (i.e., beaver complex normal conditions, beaver complex dry conditions,
171 beaver complex wet conditions, reference site normal conditions, reference site dry conditions, and
172 reference site wet conditions).

173

174 **Data analysis**

175 We calculated: (1) the annual peak, minimum, and range of NDVI in the complex and reference sites and
176 (2) rates of change in NDVI for each site type the shared months of green-up and dry down to explore how
177 beaver complexes influence riparian greenness in response to whiplash weather. We tested the influence of
178 beaver complexes on the average NDVI and annual NDVI range, peak, and minimum within a given
179 precipitation period using a repeated measures mixed effect model with complex status (beaver-influenced
180 or reference) and year treated as fixed factors and site as a random effect. We tested how peak and minimum
181 NDVI change across the transition from normal to dry to wet conditions by subsetting the data by site status
182 (complex and reference), treating climate period as a fixed factor and unique site ID (3 beaver complexes,
183 5 references) as a random effect, and using estimated marginal means and a Tukey HSD correction to
184 control for multiple comparisons across the three climate periods.

185 We distinguished green-up months as the periods between the minimum mean NDVI value and the
186 peak, whereas dry-down months were between the peak and the minimum. We assessed the influence of

187 beaver complexes on the rate of riparian green-up and dry-down during an average precipitation period
188 followed by whiplash drought to flood conditions using a repeated measures mixed effect model to test the
189 effects of complex status (beaver-influenced or reference) and time (number of days elapsed from the
190 earliest date within a given year), with site treated as a random effect. Year was included as a fixed factor
191 only for the normal and drought periods, which encompass multiple years (whereas the flooding and post-
192 flood span included the end of December 2022 and was otherwise entirely the year 2023).

193 We completed all statistical modeling in R Studio v 2023.3.0.386 using R v 4.2.3 using the
194 packages: ggplot2 (3.5.0), dplyr (1.1.3), tidyverse (2.0.0), lubridate (1.9.3), lme4 (1.1-35.1), and lmerTest
195 (0.9-40) (Grolemund and Wickham 2011; Bates et al. 2015; Wickham 2016; Kuznetsova et al. 2017;
196 Wickham et al. 2019, 2023), with coding assistance from ChatGPT (Data S4). We reported all results as
197 mean value \pm standard error and designated statistical significance at $\alpha < 0.05$.

198 **Results**

199 **Impacts of beaver influence on annual NDVI patterns across a whiplash weather period**

200 Beaver complex sites maintained a higher mean monthly NDVI relative to non-beaver influenced sites
201 across the study period (Figure 3). In the normal precipitation period (July 2017 – December 2020), beaver
202 complexes had a greater range between annual peak and minimum NDVI relative to reference sites (0.115
203 ± 0.006 vs 0.055 ± 0.004 , $F = 53.6$, $p = 0.0003$). This effect was driven by the beaver complexes maintaining
204 a higher maximum NDVI than the reference areas (0.238 ± 0.006 vs 0.149 ± 0.004 , $F = 701.1$, $p = 0.0002$),
205 while also maintaining a greater average minimum NDVI (0.123 ± 0.003 vs 0.094 ± 0.002 , $F = 81.4$, $p =$
206 0.0001). Beaver influenced areas maintained a greater range between annual peak and minimum NDVI
207 relative to the reference during drought (January 2021 - November 2022; 0.148 ± 0.008 vs 0.05 ± 0.006 ,
208 $F = 86.8$, $p < 0.0001$), with the beaver complex areas attaining a higher maximum NDVI than the reference
209 areas (0.25 ± 0.008 vs 0.13 ± 0.003 , $F = 115.8$, $p < 0.0001$), while complex and reference area minimum
210 NDVI was comparable (0.103 ± 0.01 vs 0.086 ± 0.005 , $F = 3.1$, $p = 0.13$). Following the December 2022 –
211 March 2023 flood events, beaver influenced areas no longer significantly differed in their annual NDVI
212 range relative to the non-beaver reference sites (0.075 ± 0.03 vs 0.05 ± 0.02 , $F = 0.46$, $p = 0.5$, $F = 0.46$, p
213 $= 0.5$). This shift in the extent of green up between complex and reference areas reflected that the peak
214 NDVI of the beaver complex areas (0.163 ± 0.03) was comparable to the reference peak (0.124 ± 0.02),
215 despite the complex NDVI minimum (0.088 ± 0.003) being higher than the reference (0.073 ± 0.002 , $F =$
216 14.9 , $p < 0.002$).

217 Peak NDVI in the reference areas was lower during drought (0.134 ± 0.006) and higher following
218 flood (0.172 ± 0.008) than the normal conditions (0.149 ± 0.005 , $F = 9.4$, $p = 0.0007$). In contrast, the
219 complex peak NDVI did not significantly differ between drought (0.251 ± 0.02) and normal conditions
220 (0.234 ± 0.01), while post-flood peak NDVI was reduced relative to the baseline (0.163 ± 0.02 , $F = 9.2$, p
221 $= 0.002$). Minimum NDVI in the reference areas was reduced both under drought conditions ($0.0856 \pm$
222 0.003) and following flood (0.0731 ± 0.003) relative to the baseline precipitation conditions ($0.0942 \pm$
223 0.002 , $F = 16.2$, $p < 0.0001$). This effect was paralleled by the complex NDVI, which also had significantly

224 reduced minimum annual NDVI during the drought period (0.1029 ± 0.006) and following flood
225 (0.0876 ± 0.002) than under normal precipitation conditions (0.1226 ± 0.005 , $F = 11.7$, $p = 0.0005$).

226

227 **Seasonal patterns in the rate of green-up and dry down in relation to drought status**

228 Beaver complex riparian areas greened-up for a longer duration and senesced later in the summer (and for
229 a shorter duration) than non-beaver sites in both the normal and dry periods (July 2017 - November 2022;
230 Figure 4). From July 2017 – December 2020, average NDVI in the beaver complex-influenced riparian
231 areas increased twice as rapidly than the reference sites during the shared green up months (monthly average
232 NDVI increase of 1.29×10^{-3} (beaver complex) vs. 6.72×10^{-4} (reference; $F = 39.2$, $p = 0.0007$). However,
233 the beaver complex NDVI declined more rapidly than the reference NDVI during the overlapping dry-down
234 phase (monthly average NDVI decline of -5.42×10^{-4} (beaver complex) vs. -2.33×10^{-4} (reference); $F =$
235 68.5 , $p = 0.0002$).

236 During the drought period (Figure 4), beaver complexes maintained a 3.8-fold higher NDVI green-
237 up relative to the reference sites (daily average NDVI increase of 1.08×10^{-3} (beaver complex) vs. $2.81 \times$
238 10^{-4} (reference; $F = 63.8$, $p = 0.0002$). Comparable to drought-free conditions, beaver complex NDVI
239 declined more rapidly during senescence yet still maintained a higher mean NDVI value overall (daily
240 NDVI decline of -6.47×10^{-4} (beaver complex) vs. -1.46×10^{-4} (reference; $F = 112.6$, $p < 0.0001$). High
241 peak flow disturbance began December 2022 and persisted until March of 2023. The post-flood green-up
242 phase began later than in pre-flood years (starting in April vs. January) and the subsequent dry-down phase
243 was shortened (likely due to increased water availability during the drier summer months). Beaver complex
244 monthly average NDVI increased twice as fast as the reference sites average NDVI during the overlapping
245 post-flood green up months (April – July, 1.07×10^{-3} (beaver complex) vs. 5.61×10^{-4} (reference); $F = 51.2$,
246 $p < 0.0001$). Across the entire shared green up period (January – July) the beaver complexes still maintained
247 a faster green-up rate relative to the reference sites (8.98×10^{-4} vs. 4.90×10^{-4} , $F = 43.5$, $p < 0.0001$). In the
248 post-flood 2023 dry season, beaver complexes monthly mean NDVI declined more rapidly than the
249 reference sites (-2.50×10^{-4} vs. -1.48×10^{-4} , $F = 202.5$, $p = 0.004$).

250 **Discussion**

251 The Arid West is experiencing increasing risk of both high heat and drought stress events as well
252 as more extensive flooding this century (Keeley and Syphard 2016; Swain et al. 2018; Higuera and
253 Abatzoglou 2021; Swain 2021). Beaver can buffer undesirable consequences of wetland loss and help to
254 increase the formation of multichannel drainages under increasing climate warming stressors (Beechie et
255 al. 2010; Pilliod et al. 2018; Wohl 2021). A Beaver Restoration Program established by the California
256 Department of Fish and Wildlife in 2023 exemplifies a shift in perception of the species as a problematic
257 and harvested resource to a keystone species for supporting critical wetland systems. This plan was
258 propelled by the evidence that beavers serve as climate change mitigators due to their ability to increase the
259 resiliency of wetland and riparian ecosystems to drought and wildfire (Hood and Bayley 2008; Fairfax and
260 Small 2018; Fesenmyer et al. 2018; Fairfax and Whittle 2020). However, observations of beaver activity
261 and impacts within the Mediterranean California ecoregion and in other areas of their fragmented
262 southwestern US range remains limited (Grudzinski et al. 2022), challenging the ability to develop sound
263 protocols for beaver reintroduction and restoration. We found that the Salinas River beaver complexes
264 demonstrated greater riparian greenness resistance to drought and resilience to flood disturbance than
265 nearby reference areas through a whiplash weather period, even in highly anthropogenically disturbed
266 riparian areas with sustained periods of extremely low waterflow.

267 Despite promising evidence that beaver activities and the complexes they form support the
268 development of ecological refugia from climate stressors and increase local water tables (Hood and Bayley
269 2008; Fairfax and Small 2018; Karran et al. 2018), there is ongoing debate on whether the resiliency beavers
270 provide to riparian areas under drought and wildfire extends to whiplash weather periods (Wohl 2021;
271 Larsen et al. 2021; Graham et al. 2022). We assessed how remotely derived NDVI in beaver complex
272 relative to non-complex areas responded to a high intensity pulse disturbance (extreme flood event)
273 following a chronic pressure (prolonged drought and heat stress) and as compared to a more average
274 precipitation period preceding the whiplash weather event. While a resilient system may undergo some
275 state change following a disturbance, it will be more readily able to rebound to its previous ecological state

276 as compared to a resistant system, which tends to be less responsive to a disturbance but also slower to
277 recover to the previous state (or may not recover) if change does occur (Lake 2013; Van Meerbeek et al.
278 2021). Beaver complexes demonstrated greater resistance to drought (by maintaining a comparable peak
279 NDVI to the normal climate period) and resilience to flood (greening twice as fast as the reference area,
280 although achieving a lower peak NDVI than during the normal climate period) relative to nearby reference
281 areas of the watershed. These patterns highlight the ability of these ecosystem engineers to confer multiple
282 ecological benefits under increasingly variable climate conditions.

283 For example, the ability of beavers to trap water supports greater net primary productivity (NPP)
284 and therefore carbon sequestration potential within beaver-influenced riparian areas. While NDVI is not a
285 surrogate for NPP estimation, multiple ecosystems demonstrate a positive correlation between NDVI and
286 NPP (Rafique et al. 2016). Higher mean NDVI values observed in the beaver influenced riparian zones will
287 likely have higher NPP—and thus increased carbon sequestration and storage potential following whiplash
288 weather events—relative to river areas that do not support a beaver population. Without beavers
289 maintaining high water levels, the capacity of a landscape to sequester and store carbon declines over time
290 (Laurel and Wohl 2019; Wohl 2021). Thus, sustained carbon sequestration following whiplash weather that
291 causes the complete loss of beaver dams is likely dependent upon local beaver reestablishment.

292 Paralleling other studies (Hood and Bayley 2008; Fairfax and Small 2018; Fesenmyer et al. 2018),
293 we found that immediate proximity to beaver wetlands promoted vegetation resistance to extreme drought,
294 with complexes supporting a 66% greater monthly average NDVI and a greater annual NDVI range than
295 neighboring reference areas during a prolonged, extreme drought period. Beaver complex riparian areas
296 also greened-up for a longer duration and dried-down later in the summer (but for a shorter duration) than
297 non-complex influenced areas under drought stress. Greater NDVI within the complex areas likely reflects
298 increased dam-driven soil water storage and distribution in the riparian corridor (Larsen et al. 2021). Beaver
299 dam complex areas also demonstrated vegetation growth resilience (but not resistance) as the system rapidly
300 transitioned out of drought, maintaining a higher green up rate that was comparable to the normal and dry
301 periods and higher minimum NDVI during the flood and post-flood period than the reference areas. This

302 pattern was maintained even when complexes were washed out during the flood and the river-scoured
303 structures were no longer visible post-flood, as reflected by a lower minimum NDVI for both complex and
304 reference sites than under normal and drought conditions. This resilience was likely driven by greater river
305 channel complexity and sediment build up along the beaver dammed areas. Beaver areas maintained a
306 greater minimum NDVI and more rapid green-up rate relative to the reference areas but reached a
307 comparable peak NDVI due to the greater green-up of the reference following the flood period.

308 North American beavers were historically abundant across the continent's waterways, with an
309 estimated population of upwards of 400 million individuals prior to European settlement (Boyle and Owens
310 2007). By the 1850s, beavers were nearly extirpated throughout their historic range due to extraordinary
311 hunting pressure for pelts. The beaver population has rebounded to 9-12 million individuals (Scamardo et
312 al., 2022); however, beavers often remain classified as a problematic species by natural resource governance
313 guidelines (often due to human-wildlife conflicts including undesirable flooding and tree loss) and subject
314 to management strategies including extermination (Miller and Yarrow 2015). As climate change intensifies,
315 beaver presence is increasingly positively perceived by natural resource managers due to the plethora of
316 ecosystem services they can provide through wetland creation (Brazier et al. 2021; Thompson et al. 2021).
317 The ability of beaver complexes to dampen riparian disturbance following whiplash weather events further
318 supports this paradigm shift from beavers as a pest to a 'natural climate solution.'

319 While we observed both NDVI resistance and resilience in response to whiplash weather patterns,
320 the extent of these responses was limited to within an approximately 200-meter lateral distance from beaver
321 dams in our study system, which is highly modified by agricultural and urban development that include
322 damming, water diversion, and off-road vehicle uses in the riparian corridor. Given these structural
323 limitations to their expansion and low density due to historic extirpation (Baker and Hill 2003; Carrillo et
324 al. 2009; White et al. 2015), the Salinas River beaver complexes still exhibited resistance to drought and
325 resilience to high peak flow events with regards to plant community greenness. Our research reinforces the
326 concept that policies promoting beaver re-colonization into fragmented habitats can help to confer both
327 riparian stability in the face of increasingly erratic climatic conditions. By supporting nearly year-round

328 green riparian areas adjacent to their dam complexes and increasing *both* riparian corridor vegetation
329 resistance to extreme, chronic drought and resiliency to extraordinary flood events, these findings showcase
330 the ability of beavers to serve as a natural climate change mitigators and lends support to policies that
331 encourage beaver relocation and restoration within the edges and fragmented patches of their historic
332 habitat ranges.

333 **Supplemental material**

334

335 **Data S1.** A python script for calculating normalized difference vegetation index (NDVI) in the study focal
336 areas from the USGS EarthExplorer website. In brief, 30-m resolution Landsat 8 and 9 Imagery (10% or
337 less cloud cover) was collected using United States Geographical Survey’s webtool Earth Explorer from
338 2017-07-01 to 2023-12-31 using Landsat C2 U.S. Analysis Ready Data. Band 4 (red) and Band 5 (near
339 infrared) .tif files were downloaded to calculate NDVI.

340

341 **Data S2.** A python notebook for deriving a zonal statistics table for normalized difference vegetation index
342 data in the Salinas River, California study region encompassing beaver complex and reference areas.

343

344 **Data S3.** Comma separated value file of remotely sensed mean normalized difference vegetation index data
345 for 3 beaver complexes and 5 non-beaver reference areas along the Salinas River, California from July 2017
346 until December 2023 derived from Landsat C2 U.S. Analysis Ready Data on the USGS EarthExplorer
347 website and Data S1 and Data S2 scripts.

348

349 **Data S4.** Statistical models used to analyze how normalized difference vegetation index in beaver-complex
350 and nearby reference areas of the Salinas River, CA respond to whiplash weather (July 2017 through
351 December 2023) using an RMD file to process the associated ‘Data S3.csv’ file.

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369

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374

375 Any use of trade, product, website, or firm names in this publication is for descriptive
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644 **Figure Legends**

645 Figure 1. A map of beaver *Castor canadensis* complex and reference (non-beaver complex) sites along the
646 Salinas River in San Luis Obispo County, California where change in remotely sensed normalized
647 difference vegetation index (NDVI) from July 2017 – December 2023 was used to assess the relative
648 resistance and resilience of beaver complex riparian systems to whiplash weather. The “BW” sites
649 represent three beaver wetlands whereas the “C” sites represent five reference sites. The NDVI value for
650 May 22, 2022 is presented and the inset shows where the study region is located within California.

651
652 Figure 2. Images taken from Fall 2022 site visits sites along the Salinas River in San Luis Obispo County,
653 California showing three beaver *Castor canadensis* complex influenced riparian areas (a–c) and three
654 nearby reference areas outside of the direct influence of the beaver complexes (d–f).

655
656 Figure 3. Monthly average normalized difference vegetation index (NDVI) distinguished by site type
657 (beaver *Castor canadensis* complex or wetland) and period (normal, drought, flood, wet) from July 2017 –
658 December 2023 along the Salinas River in San Luis Obispo County, California. The center line denotes
659 the median value, while the box contains the 25th to 75th percentiles, and whiskers mark the 5th and 95th
660 percentiles. Significant differences in average NDVI between the beaver complex and reference area for
661 each period are denoted (***) $p < 0.001$, (**) $p < 0.01$, (*) $p < 0.05$.

662
663 Figure 4. Monthly average normalized difference vegetation index (NDVI) in beaver *Castor canadensis*
664 complex and reference (non-beaver complex) sites along the Salinas River in San Luis Obispo County,
665 California over a whiplash weather period (July 2017 – December 2023). The vertical bars represent the
666 standard deviation over the monthly sample periods and the shaded areas surrounding the lines highlight
667 the months in which sites experience a green-up (i.e., an increase in mean NDVI relative to the minimum
668 in a given year). Beaver complex green up and senescence rates were significantly greater than reference
669 areas in all the weather periods ($p < 0.005$ in all cases).